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Tanya L. Alderete, University of Colorado Boulder
Ran Jin, University of Southern California
Douglas Walker, Emory University
Damaskini Valvi, Harvard University
Zhanghua Chen, University of Southern California
Dean P Jones, Emory University
Cheng Peng, University of Southern California
Frank D. Gilliland, University of Southern California
Kiros Berhane, University of Southern California
David V. Conti, University of Southern California

Only first 10 authors above; see publication for full author list.

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Perfluoroalkyl substances, metabolomic profiling, and alterations in glucose homeostasis among overweight and obese Hispanic children: A proof-of-concept analysis

Tanya L. Alderete a,1, Ran Jin b,1, Douglas I. Walker c,1, Damaskini Valvi d, Zhanghua Chen b, Dean P. Jones e, Cheng Peng b, Frank D. Gilliland b, Kiros Berhane b, David V. Conti b, Michael I. Goran f, Lida Chatzi b,⁎

a Department of Integrative Physiology, University of Colorado Boulder, Boulder, CO, United States
b Department of Preventive Medicine, University of Southern California, Los Angeles, CA, United States
c Department of Environmental Medicine and Public Health, Icahn School of Medicine at Mount Sinai, New York, NY, United States
d Department of Environmental Health, Harvard T.H. Chan School of Public Health, Boston, MA, United States
e Clinical Biomarkers Laboratory, Division of Pulmonary, Allergy, Critical Care and Sleep Medicine, Emory University, Atlanta, GA, United States
f Department of Pediatrics, Children’s Hospital of Los Angeles, The Saban Research Institute, United States

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ABSTRACT

Objective: To examine the prospective associations between exposure to perfluoroalkyl substances (PFASs) and longitudinal measurements of glucose metabolism in high-risk overweight and obese Hispanic children.

Methods: Forty overweight and obese Hispanic children (8–14 years) from urban Los Angeles underwent clinical measures and 2-hour oral glucose tolerance tests (OGTT) at baseline and a follow-up visit (range: 1–3 years after enrollment). Baseline plasma perfluorooctanoic acid (PFOA), perfluorooctane sulfonate (PFOS), perfluorohexane sulfonic acid (PFHxS), and the plasma metabolome were measured by liquid-chromatography with high-resolution mass spectrometry. Multiple linear regression models were used to assess the association between baseline PFASs and changes in glucose homeostasis over follow-up. A metabolome-wide association study coupled with pathway enrichment analysis was performed to evaluate metabolic dysregulation associated with plasma PFASs concentrations. We performed a structural integrated analysis aiming to characterize the joint impact of all factors and to identify latent clusters of children with alterations in glucose homeostasis, based on their exposure and metabolomics profile.

Results: Each ln (ng/ml) increase in PFOA and PFHxS concentrations was associated with a 30.6 mg/dL (95% CI: 8.8–52.4) and 10.2 mg/dL (95% CI: 2.7–17.7) increase in 2-hour glucose levels, respectively. A ln (ng/ml) increase in PFHxS concentrations was also associated with 17.8 mg/dL increase in the glucose area under the curve (95% CI: 1.5–34.1). Pathway enrichment analysis showed significant alterations of lipids (e.g., glycosphingolipids, linoleic acid, and de novo lipogenesis), and amino acids (e.g., aspartate and asparagine, tyrosine, arginine and proline) in association to PFASs exposure. The integrated analysis identified a cluster of children with increased 2-h glucose levels over follow up, characterized by increased PFAS levels and altered metabolite patterns.

Conclusions: This proof-of-concept analysis shows that higher PFAS exposure was associated with dysregulation of several lipid and amino acid pathways and longitudinal alterations in glucose homeostasis in Hispanic youth. Larger studies are needed to confirm these findings and fully elucidate the underlying biological mechanisms.

Abbreviations: PFASs, perfluoroalkyl substances; PFHxS, perfluorohexane sulfonic acid; PFOA, perfluorooctanoic acid; PFOS, perfluorooctane sulfonate; EDCs, endocrine-disrupting chemicals; PPARs, peroxisome proliferator activated receptors; OGTT, oral glucose tolerance test; HOMA-IR, homeostatic model assessment for insulin resistance; AUC, area under the curve; BMI, body mass index; DEXA, dual-energy X-ray absorptiometry; FDR, false discovery rate; HRM, high-resolution metabolomics

⁎ Corresponding author at: 2001 North Soto Street, Los Angeles, CA 90089-9239, United States.
E-mail addresses: tanya.alderete@colorado.edu (T.L. Alderete), jinr@usc.edu (R. Jin), douglas.walker@mssm.edu (D.I. Walker), dvalvi@hsph.harvard.edu (D. Valvi), zhanghuc@usc.edu (Z. Chen), djpjones@emory.edu (D.P. Jones), chengpen@usc.edu (C. Peng), gillilan@usc.edu (F.D. Gilliland), kiros@usc.edu (K. Berhane), dconti@med.usc.edu (D.V. Conti), goran@usc.edu (M.I. Goran), chatzi@usc.edu (L. Chatzi).

1 Authors contributed equally.
1. Introduction

Type 2 diabetes is a silent epidemic in youth and its incidence has continuously increased over the past 2 decades, with the fastest rise observed among Hispanics compared to non-Hispanic whites (Lascar et al., 2018; Mayer-Davis et al., 2017). Young-onset type 2 diabetes has a more aggressive disease phenotype that can lead to premature development of complications and long-term adverse health effects with direct impacts on quality of life (Lascar et al., 2018). A growing body of evidence indicates that early life environmental exposures can result in metabolic abnormalities and increased type 2 diabetes risk in later life (Alderete et al., 2017; Alonso-Magdalena et al., 2011; Heindel et al., 2011), yet there has been little study of the environmental contributions to diabetes risk in minority ethnic groups.

Perfluoroalkyl substances (PFASs) include chemicals that have been used for decades as industrial surfactants and in textile coatings, firefighting foams and consumer products (e.g., cookware, food containers, and clothing) (Grandjean and Clapp, 2014). Rodent studies suggest that perinatal and/or early postnatal exposure to PFOA, PFOS and PFHxS induces increased insulin levels and impaired glucose tolerance in offspring (Hines et al., 2009; Wan et al., 2014; Lv et al., 2013; Zhao et al., 2011). Cross-sectional epidemiological studies in adults from the National Health and Nutrition Examination Survey (NHANES) support observations from animal studies, showing that increased serum PFOA and PFOS levels were associated with increased fasting glucose (Liu et al., 2018), increased insulin levels, and insulin resistance (indicated by increased homeostatic model assessment for insulin resistance, HOMA-IR) (Lin et al., 2009). Further, PFOA serum levels was positively associated with diabetes prevalence in US men in the NHANES study (He et al., 2017), and a higher incidence of type 2 diabetes in women in the Nurses’ Health Study II (He et al., 2017). Few previous studies assessed associations between PFAS exposures with diabetes risk in children. The National Health and Nutrition Examination Survey (NHANES) study suggested that increased serum perfluorononanoic acid (PFNA) concentrations were associated with decreased β-cell function and increased risk for hyperglycemia in US adolescents (aged 12–20 years) (Lin et al., 2009), while the European Youth Heart Study showed that childhood PFOA exposure at 9 years of age was associated with decreased β-cell function at 15 years of age (Domaset al., 2016). However, a pregnancy cohort in the US showed negative associations between plasma PFAS concentrations and insulin resistance at the age of 8 years (Fleisch et al., 2017).

The mechanisms underlying the effects of PFAS exposure on dysregulated glucose metabolism and diabetes risk remain unknown. In a high-throughput targeted metabolomics study, PFOA exposure has been associated with metabolic disruption in liver tissue of mice, such as alterations in lipid metabolism (e.g., glycerophospholipids, linoleic acid, and arachidonic acid) and amino acids (e.g., tyrosine, tryptophan, arginine and proline) (Yu et al., 2016). To our knowledge, only one human study has examined PFAS exposure and the plasma metabolome showing that circulating levels of PFASs in elderly adults was associated with dysregulated glycerophosphocholines and fatty acids metabolism (Salihovic et al., 2019). Dysregulation of lipid and amino acid pathways have been well-characterized and found to be strongly associated with the risk of developing type 2 diabetes in human studies (Guasch-Ferre et al., 2016; Padberg et al., 2014). Collectively, results from these studies indicate that PFAS exposure might cause metabolic perturbations in lipid and amino acid pathways, thereby contributing to increased risk for type 2 diabetes.

The objective of this study was to examine the associations between PFASs exposure and longitudinal measurements of glucose metabolism in overweight and obese Hispanic children. We also aimed to understand the underlying metabolic disturbances due to PFASs exposure by performing a high-resolution metabolomics analysis. We hypothesized that higher PFASs exposures would be associated with dysregulated glucose homeostasis and alterations in key metabolic pathways implicated in type 2 diabetes pathophysiology.

2. Methods

2.1. Study population

We examined participants from the Study of Latino Adolescents at Risk of Type 2 Diabetes (SOLAR) project, which is a longitudinal cohort that recruited participants in two waves from 2001 to 2012 (Goran et al., 2004; Weigensberg et al., 2003). Participants lived in Los Angeles, California, and were recruited predominantly from metabolic clinics as well as word of mouth, health fairs, and advertisements in the local communities. Inclusion criteria were Hispanic/Latino ethnicity (defined as self-reported race/ethnicity for all participants, parents, and grandparents), age- and sex-specific BMI equal or above the sex- and age-specific 85th percentile of the Centers for Disease Control and Prevention (CDC) growth charts, and absence of type 1 or type 2 diabetes. Participants were also excluded if they were using a medication or diagnosed with a condition known to influence insulin and/or glucose metabolism or body composition. Participants underwent repeated, detailed phenotyping for clinical risk factors of type 2 diabetes using as well as body composition testing. The original sample size from the full cohort was 258 children with the necessary 2-hour OGTT outcome measures, covariate data, and plasma samples. Of these 258 children, we chose a random sample of 40 for this proof-of-concept study. These 40 participants did not significantly differ from the 258 based on important characteristics such as age, body fat percent, or measures from the 2-hour OGTT (Supplementary Table 1). These participants had their baseline visit between the 2001–2011 and were followed for an average of 1.3 years (standard deviation: 0.5). Study protocol was approved by the University of Southern California (USC) Institutional Review Board (IRB) and informed written consent (parental consent for participants < 18 years) and assent (when applicable) were obtained for each participant before initiation of the study.

2.2. Clinical assessments

Participants attended annual clinical visits at the Los Angeles County Children’s Hospital or the USC University Hospital. At the visits, participants received a comprehensive medical history and physical examination and pubertal staging was determined using the Tanner scale (Marshall and Tanner, 1969; Marshall and Tanner, 1970). Height (m) and weight (kg) were measured to determine body mass index (BMI) and the age- and sex-specific CDC growth charts were used to classify BMI status (Kuczmarski et al., 2000). A DEXA scan was also performed to determine body fat percent using a Hologic QDR 4500 W (Hologic, Bedford, MA). A 2-hour OGTT was performed at each clinical visit using a glucose load of 1.75 g/kg body weight, to a maximum of 75 g, of anhydrous glucose dissolved in water. Baseline and post-challenge samples were assayed for glucose and insulin at fasting as well as 30, 60, and 120 min after glucose intake and were used to calculate glucose and insulin areas under the curve (AUC).The homeostatic model assessment (HOMA-IR), an estimate of insulin resistance, was calculated using fasting glucose and insulin values from the 2-hour OGTT by using the formula: HOMA-IR = fasting glucose (mg/dL) × fasting insulin (μU/mL) / 405 (Matthews et al., 1985). Glucose was assayed using a Yellow Springs Instruments analyzer (YSI INC. Yellow Springs, OH, USA) that uses a membrane bound glucose oxidase technique. Lastly, insulin was assayed using an automated enzyme immunoassay ( Tosoh AIA 600 II analyzer, Tosoh Bioscience, Inc., South San Francisco, CA, USA).

2.3. High-resolution metabolomics

High resolution metabolomics (HRM) profiling was completed using standardized methods (Soltow et al., 2013). Samples were prepared and
analyzed in a single batch and included six analyses of pooled human plasma (CHEAR-Ref) for quality control purposes and reference standardization. Plasma aliquots were removed from storage at −80 °C, thawed and 50 μL was treated with 100 μL of ice-cold LC-MS grade acetonitrile. Plasma was then equilibrated for 30 min on ice, centrifuged (16.1 × g at 4 °C) for 10 min to remove precipitated proteins, transferred to a 200 μL autosampler vial and maintained at 4 °C until analysis (< 22 h). Sample extracts were analyzed using liquid chromatography and Fourier transform high-resolution mass spectrometry (Dionex Ultimate 3000, Q-Exactive HF, Thermo Scientific). The chromatography system was operated in a dual pump configuration that enabled parallel analyte separation and column flushing. For each sample, 10 μL aliquots were analyzed in triplicate using hydrophilic interaction liquid chromatography (HILIC) with electrospray ionization (ESI) source operated in positive mode for metabolic profiling and reverse phase chromatography with ESI operated in negative mode for quantification of PFASs. Analyte separation for HILIC was accomplished by a 2.1 mm × 50 mm × 2.5 μm Waters XBridge BEH Amide XP HILIC reverse phase chromatography with ESI operated in negative mode for metabolomic profiling and reverse phase chromatography with ESI operated in negative mode for quantification of PFASs. Analyte separation for HILIC was accomplished by a 2.1 mm × 50 mm × 2.5 μm Waters XBridge BEH Amide XP HILIC and an eluent gradient (A = 2% formic acid, B = water, C = acetonitrile) consisting of an initial 1.5 min period of 2.5% A, 22.5% B, 75% C followed by a linear increase to 2.5% A, 77.5% B, 20% C at 4 min and a final hold of 1 min. RPC separation was by 2.1 mm × 50 mm × 3 μm endcapped C18 column (Higgins) using an eluent gradient (A = 2% 5 mM ammonium acetate, B = water, C = acetonitrile) consisting of an initial 2 min period of 5%, 90%, 5% C, followed by a linear increase to 5%A, 0%B, 95%C at 6 min and held for the remaining 4 min. For both methods, mobile phase flow rate was held at 0.35 mL/min for the first 1.5 min, increased to 0.5 mL/min and held for the final 4 min. The high-resolution mass spectrometer was operated at 120,000 resolution and mass-to-charge ratio (m/z) range 85–2175. Probe temperature, capillary temperature, sweep gas and S-Lens RF levels were maintained at 200 °C, 300 °C, 1 arbitrary units (AU), and 45, respectively, for both polarities. Additional source tune settings were optimized for sensitivity using a standard mixture, positive tune settings for sheath gas, auxiliary gas, sweep gas and spray voltage setting were 45 AU, 25 AU and 3.5 kV, respectively; negative settings were 30 AU, 5 AU and −3.0 kV. Maximum C-trap injection times of 100 milliseconds and automatic gain control target of 1 × 10^8 for both polarities. During untargeted data acquisition, no exclusion or inclusion masses were selected, and data was acquired in MS1 mode only. Raw data files were then extracted using aplCMS (Yu et al., 2009) with modifications by xMSanalyzer (Uppal et al., 2013). Uniquely detected ions consisted of m/z, retention time and ion abundance, referred to as m/z features. Prior to data analysis, m/z features were batch corrected using CombBat (Johnson et al., 2007) and filtered to remove those with coefficient of variation (CV) ≥100% and > 10% non-detected values.

2.4. Quantification of plasma levels of PFASs

Concentrations of PFOA, PFOS and PFHxS were quantified by reference standardization using the LC-HRMS method described in Section 2.3 with reverse phase chromatography for analyte separation and negative mode ESI (Go et al., 2015). Analyte identification was confirmed by matching MS² ion dissociation patterns, precursor m/z and retention time to authentic reference standards, and concentrations in the Children’s Health Exposure Analysis Resource (CHEAR) pooled plasma reference sample were quantified by methods of addition and comparison against NIST standard reference material 1950 (Metabolites in Frozen Human Plasma) (Simon-Manso et al., 2013). Using the CHEAR reference samples, the response factor for each analyte was determined using the M-H adduct, and plasma concentrations were calculated in study samples by single point calibration via response factors (calculated as the ratio between the known concentration of the compound being quantified and ion intensity in CHEAR reference samples. Calculated limit of detection (LOD) for PFOA, PFOS and PFHxS was 0.02, 0.1 and 0.03 ng/mL, respectively. PFOA, PFOS, and PFHxS were detected in 97.5%, 97.5%, and 100% of participants. Due to the moderate to high correlation of PFAS (r from 0.4 to 0.7, Table 2), we also performed a principal component (PC) analysis for the 3 PFAS (PFOA, PFOS, PFHxS), and we selected the first component (“PC1”), as a composite variable representing PFAS burden, which explained 96.7% of the variance. This variable was used as the primary exposure variable in the integrated analysis with the metabolites and the outcomes of interest.

2.5. Statistical analysis

Geometric means (GMs), and interquartile ranges were calculated for plasma concentrations of all PFASs. Additionally, Spearman correlation coefficients among PFAS concentrations were calculated. Since the distribution of PFASs were right-skewed, exposures were natural log transformed for statistical analyses. Changes in metabolic outcomes were calculated as the respective follow-up measure subtracted from the baseline measure (e.g., fasting glucose at follow-up − baseline fasting glucose). We fitted generalized additive models (GAM) with penalized spline smooth terms and visually assessed plotted splines to determine linearity of exposure–outcome associations. We found no evidence of nonlinear associations of PFAS concentrations with type 2 diabetes outcomes except for the association of PFHxS concentrations with changes in the glucose AUC (pGAM < 0.05). We therefore fitted multiple linear regression models to estimate the relationships between change in metabolic outcomes (i.e., fasting glucose and insulin, glucose and insulin AUC, HOMA-IR) in relation to each PFAS exposure, as continuous natural log–transformed plasma concentrations at baseline. For glucose AUC and PFHxS, cubic spline models were also fitted to investigate the non-linear association. Baseline pubertal status was classified as pre-puberty (Tanner Stage 1), puberty (Tanner Stage 2–4), and post-puberty (Tanner Stage 5). A modified version of the Hollingshead Four-Factor Index of Social Status was used to assess socioeconomic status in participants where information was available (n = 33). This index takes into account the occupation and education of each parent/guardian residing in the child’s home in order to generate a single measure of a family’s social status. Social position was then categorized as < 25th percentile (n = 10), > 25th percentile and < 75th percentile (n = 19), ≥75th percentile (n = 4), and missing (n = 7). Multiple linear models adjusted for the baseline outcome, baseline social position, age, sex, and change in age at follow-up (Model 1). Models were further adjusted for baseline pubertal status as well as baseline and change in body fat percent at follow-up as adiposity and puberty are strong predictors of the outcome of interest (Model 2). Selections of covariates retained in the models were based on directed acyclic graphs (Supplementary Fig. 1) (Textor et al., 2011). Prior studies have shown more pronounced associations between PFASs exposure and insulin resistance in females (Halldorsson et al., 2012), therefore effect modification by sex was examined via an interaction term and stratification. As an exploratory analysis, effect modification of obesity (≥98th versus < 98th CDC percentile) and puberty (Tanner 1 versus Tanner 2–5) were also examined.

The metabolome-wide association study was performed in order to identify metabolites and global metabolic changes associated with PFASs exposure, including PFOA, PFOS and PFHxS. This untargeted analysis fitted multiple linear models that were used to examine the associations between plasma PFASs concentrations (independent variables) and the log_2 transformed m/z features (dependent variables) after controlling for baseline age, sex, and social position. The log_2 transformation of metabolite data was used to meet the assumptions of linear regression. The m/z features with a Benjamini-Hochberg false discovery rate (FDR) of ≤ 20% (Benjamini and Hochberg, 1995) were then selected for visualization by Manhattan plots and metabolic pathway enrichment analysis using Mummichog (Li et al., 2013) with 10,000 permutations. Using this pathway enrichment analysis, we then identified significantly dysregulated pathways associated with PFASs.
exposure (p ≤ 0.05).

We performed an integrated latent variable analysis to characterize the joint impact of all factors and identify latent clusters of children with alterations in glucose homeostasis, based on their PFAS exposure and metabolites profile by LUCIDus R package (https://CRAN.R-project.org/package=LUCIDus). For the estimation of the number of latent clusters, we used Bayesian Information Criteria. We used the PFASs composite variable as the primary exposure variable instead of the three individual PFASs since highly correlated exposure parameters invalidate the presumption in this analysis. Metabolites were annotated using adduct of protonated ion (M + H) (Supplementary Table 2) and were selected from the most significantly altered metabolic pathways (based on p value and the numbers of significant metabolite features) associated with the plasma PFAS in the Mummichog analysis. The outcome variable was determined as the change in 2-hour post-load glucose levels from baseline to follow-up visit (binary: increase vs. decrease). The LUCIDus package provides effect estimates for the association of estimated latent clusters with changes in 2 h glucose levels. It also provides the distribution of exposure and metabolites in each identified cluster (assigned based on the posterior probability using cutoff point of 0.05) (Supplementary Table 3).

All statistical analyses were performed using SAS, version 9.4 (SAS, Institute, Cary, NC) and the R statistical environment version 3.1.2.

3. Results

3.1. Characteristics of the study population

Overweight and obese Hispanic children (BMI percentile, mean ± SD: 96.8 ± 3.5) included in this study were between 8 and 14 years of age at baseline (Table 1). Half of children were female, and most were in the early to mid-stages of puberty. On average, children had normal fasting (<100 mg/dL) and 2-hour glucose levels (<140 mg/dL) at baseline. Mean fasting insulin levels were 16.1 mg/dL (SD 17.1) and mean HOMA-IR was 3.5 (SD 2.5). Fasting glucose levels were significantly higher in the follow up clinical visit compared to baseline (mean: 89.4 mg/dL vs. 97.7 mg/dL, p < 0.05, Table 1).

3.2. Plasma PFASs were associated with clinical risk factors for type 2 diabetes

PFOS, PFOA, and PFHxS were detected in 97.5% of participants and geometric means were 12.22, 2.78, and 1.65 ng/mL, respectively (Table 2). PFASs concentrations showed moderate to high pairwise correlations, with the stronger correlation seen between PFOA and PFOS (Spearman r = 0.71) (Table 2). As shown in Table 3 and Fig. 1, for each ln-unit increase in PFOA and PFHxS concentrations, there was 30.6 mg/dL (95% CI: 8.8–52.4) and 10.2 mg/dL (95% CI: 2.7–17.7) increase in 2-hour glucose levels between the baseline and follow-up visit, respectively. Additionally, for each ln-unit increase in PFHxS concentrations, there was a 17.8 mg/dL increase in the glucose AUC (95% CI: 1.5–34.1). There was no evidence for effect modification by sex; however, there was evidence for effect modification by pubertal status when examining the relationships between PFHxS and glucose AUC (p interaction = 0.03). For each ln-unit increase in PFHxS concentrations, there was a 46 mg/dL (95% CI: 16.3–75.5) increase in the glucose AUC among children who were in puberty or post-puberty (Tanner Stage 2–5) while this relationship was not observed among pre-pubertal children (Tanner Stage 1) (β = 63.8; 95% CI: −59.2–186.9). Lastly, there was no effect modification by BMI status (p interactions ≥ 0.2) and PFASs exposures were not significantly associated with other metabolic outcomes (i.e., fasting glucose, fasting insulin, and insulin AUC, HOMA-IR).

3.3. PFASs exposure was associated with plasma metabolites and metabolic pathways

To identify metabolic alterations associated with PFASs, we performed the metabolome-wide association study and identified 149, 298, and 17 metabolite features associated with plasma concentrations of PFOA, PFOS and PFHxS, respectively, at FDR < 20% (Fig. 2). We next performed the Mummichog pathway enrichment analysis with the input of all detected metabolite features correlated with PFOA, PFOS and PFHxS, and identified 24 metabolic pathways that were associated with PFASs exposure (Fig. 3). Exposure to PFASs was associated with dys-regulation of multiple lipid metabolic pathways that included glycosphingolipid metabolism, fatty acid metabolism, de novo lipogenesis, and linoleic acid metabolism (Fig. 3). A series of amino acid metabolic pathways were also associated with PFASs exposure, including aspartate and asparagine, tyrosine, and arginine and proline metabolism (Fig. 3). Other significantly altered metabolic pathways included amino sugar metabolism, vitamins and cofactors (e.g., vitamin B3 and 9), as well as nitrogen metabolism.

3.4. Identification of subgroups of children with alterations in glucose homeostasis using an integrated latent variable analysis

Two latent clusters were identified. “Cluster 2” was associated with increased 2-h glucose levels between the baseline and follow-up visit (Fig. 4, red line connecting “Cluster 2” and the outcome). The OR for increased 2-h glucose levels at follow up associated with this cluster was 8.63 (compared to “Cluster 1”). This “high-risk” cluster was also positively associated with the PFAS composite variable (blue line connecting PFAS to “Cluster 2”) and with an altered plasma metabolite pattern. This pattern was characterized by increased plasma levels (blue lines connecting “Cluster 2” to metabolites) of palmitic acid (de novo lipogenesis pathway), hydroxyproline (HPODE, linoleate metabolism), tyrosine and phenylalanine (tyrosine metabolism), and arginine (arginine and proline metabolism). It was also characterized by decreased plasma levels (grey lines connecting “Cluster 2” to metabolites) of sphingomyelin (glycosphingolipid pathway), linoleic acid (linoleate metabolism) and aspartate (aspartate and asparagine metabolism) (Fig. 4 and Supplementary Table 3). In order to characterize these clusters qualitatively, we assigned each child to one of two clusters based on an estimated probability > 0.5 for membership within a cluster. Children assigned to the “high-risk” cluster (i.e. “Cluster 2”) had higher PFAS concentrations, associations reflective of the metabolites of previously mentioned to define “Cluster 2”, and a substantially increased risk for higher 2h glucose levels than children in
Further adjustment for pubertal status (categorical) as well as baseline and change in body fat percent at follow-up.

Model 1: adjusted for sex, baseline social position (categorical), baseline outcome as well as baseline and change in age at follow-up. Model 2: Model 1 covariates plus Hispanic youth in the current study. However, a recent pregnancy at age 15 years. Plasma concentrations of PFOS and PFOA in the Danish subjects of the NHANES were comparable to the average serum concentrations quantified for PFOS, PFOA and PFHxS in adolescents and/or Mexican Americans in NHANES between 1999 and 2010, and almost twice as high as the PFASs concentrations reported for the same NHANES subpopulation between 2011 and 2014, which is likely due to the decreasing environmental levels of these substances in U.S. and elsewhere (Fourth National Report on Human Exposure to Environmental Chemicals, 2017).

In the current study, PFASs exposure was only associated with the change in 2-hour glucose levels and the change in the glucose AUC but not other metabolic outcomes (e.g., change in fasting glucose and insulin levels, HOMA-IR). This may be due to the fact that the current study included otherwise healthy overweight and obese Hispanic youth. Further, it is possible that euglycemia in the fasting state was maintained despite important physiological changes in insulin resistance and β-cell function that may only be observed following a glucose challenge. Additionally, the consequences of insulin resistance may be more apparent after an oral glucose challenge when greater insulin levels are needed to maintain normoglycemia compared to the fasting state (Gerich, 2003). We also found that higher PFHxS concentrations were associated with an increase in the glucose AUC among children who were in puberty or post-puberty. Work by our group has shown reductions in insulin sensitivity and β-cell function during the pubertal transition (Goran and Gower, 2001). Therefore, any declines in insulin sensitivity and/or β-cell function that resulted from PFASs exposure may have been exacerbated by the normal declines that are seen during this transition. It is possible that discrepancies between findings in this study and prior work in adolescents (Lin et al., 2009; Domazet et al., 2016) may be due to failing to account for pubertal status.

The exact mechanisms underlying the associations between PFASs exposure and diabetes-related traits remain uncertain. Rapid advancement in metabolomics technologies provides us a unique tool to examine the metabolic perturbations in response to environmental exposures (Salihovic et al., 2016; Andrianou et al., 2017). Consistent with another human study on PFAS (Salihovic et al., 2019), and other studies examining persistent organic pollutants (e.g. organochlorine pesticides) and alterations in plasma metabolome (Salihovic et al., 2016), we found that plasma PFAS concentrations were significantly associated with dysregulation of multiple lipid pathways (e.g., metabolism of

Table 2
Distribution of polyfluoroalkyl substances (PFASs) concentrations and Spearman correlation coefficients for PFASs measured at baseline among overweight and obese Hispanic participants living in urban Los Angeles, CA who had their baseline visits between 2001 and 2011.

<table>
<thead>
<tr>
<th>Exposure</th>
<th>PFHxS</th>
<th>PFOS</th>
<th>PFOA</th>
</tr>
</thead>
<tbody>
<tr>
<td>Plasma concentration of PFASs (ng/ml)</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Geometric mean</td>
<td>1.65</td>
<td>12.22</td>
<td>2.78</td>
</tr>
<tr>
<td>Geometric mean SD</td>
<td>2</td>
<td>1.91</td>
<td>1.29</td>
</tr>
<tr>
<td>Minimum</td>
<td>0.47</td>
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<td>1.88</td>
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<tr>
<td>Maximum</td>
<td>12.81</td>
<td>65.3</td>
<td>5.37</td>
</tr>
<tr>
<td>Below LOD (%)</td>
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<td>0</td>
<td>2.5</td>
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</table>

Spearman correlation coefficients

<table>
<thead>
<tr>
<th>PFHxS</th>
<th>PFOS</th>
<th>PFOA</th>
</tr>
</thead>
<tbody>
<tr>
<td>1.00</td>
<td></td>
<td></td>
</tr>
<tr>
<td>0.58</td>
<td>1.00</td>
<td></td>
</tr>
<tr>
<td>0.44</td>
<td>0.71</td>
<td>1.00</td>
</tr>
</tbody>
</table>

*p-Value < 0.005.

**p-Value < 0.0001.

Cluster 1 (Supplemental Table 3).

4. Discussion

This is the first prospective study in Hispanic children to study environmental chemical exposures and alterations in glucose homeostasis using longitudinal clinical measures of insulin and glucose metabolism. We found that higher plasma PFASs concentrations were associated with dysregulated glucose levels following an oral glucose tolerance test. We also combined, for the first time, plasma PFASs chemical analysis and untargeted metabolomics analyses showing that PFAS exposure may be associated with changes in key metabolic pathways underlying type 2 diabetes pathophysiology such as metabolism of glycosphingolipids, fatty acid, linoleic acid, tyrosine, aspartate and asparagine, arginine and proline.

Several prospective studies in adults have shown that PFASs exposure may be associated with increased risk for type 2 diabetes (Halldorsson et al., 2012; Sun et al., 2018; Matilla-Santander et al., 2017). While longitudinal studies in children are scarce, a previous prospective study in Danish children (Domazet et al., 2016) found an association between PFOA exposure and decreased beta-cell function at age 15 years. Plasma concentrations of PFOS and PFOA in the Danish children were 3 times higher than the concentrations observed in the Hispanic youth in the current study. However, a recent pregnancy cohort study in the U.S. found that children with higher PFASs exposure had lower insulin resistance (Fleisch et al., 2017). The discrepancies in results among studies may be due to differences in study design, population demographics, or PFASs concentrations distributions. However, in the current study, PFASs concentrations were comparable to the average serum concentrations quantified for PFOS, PFOA and PFHxS in adolescents and/or Mexican Americans in NHANES between 1999 and 2010, and almost twice as high as the PFASs concentrations reported for the same NHANES subpopulation between 2011 and 2014, which is likely due to the decreasing environmental levels of these substances in U.S. and elsewhere (Fourth National Report on Human Exposure to Environmental Chemicals, 2017).

In the current study, PFASs exposure was only associated with the change in 2-hour glucose levels and the change in the glucose AUC but no other metabolic outcomes (e.g., change in fasting glucose and insulin levels, HOMA-IR). This may be due to the fact that the current study included otherwise healthy overweight and obese Hispanic youth. Further, it is possible that euglycemia in the fasting state was maintained despite important physiological changes in insulin resistance and β-cell function that may only be observed following a glucose challenge. Additionally, the consequences of insulin resistance may be more apparent after an oral glucose challenge when greater insulin levels are needed to maintain normoglycemia compared to the fasting state (Gerich, 2003). We also found that higher PFHxS concentrations were associated with an increase in the glucose AUC among children who were in puberty or post-puberty. Work by our group has shown reductions in insulin sensitivity and β-cell function during the pubertal transition (Goran and Gower, 2001). Therefore, any declines in insulin sensitivity and/or β-cell function that resulted from PFASs exposure may have been exacerbated by the normal declines that are seen during this transition. It is possible that discrepancies between findings in this study and prior work in adolescents (Lin et al., 2009; Domazet et al., 2016) may be due to failing to account for pubertal status.

The exact mechanisms underlying the associations between PFASs and diabetes-related traits remain uncertain. Rapid advancement in metabolomics technologies provides us a unique tool to examine the metabolic perturbations in response to environmental exposures (Salihovic et al., 2016; Andrianou et al., 2017). Consistent with another human study on PFAS (Salihovic et al., 2019), and other studies examining persistent organic pollutants (e.g. organochlorine pesticides) and alterations in plasma metabolome (Salihovic et al., 2016), we found that plasma PFAS concentrations were significantly associated with dysregulation of multiple lipid pathways (e.g., metabolism of
glycosphingolipid, fatty acid, and linoleic acid). Importantly, alteration of such lipid metabolites was also reported to be associated with type 2 diabetes risk in metabolomics studies, as reviewed elsewhere (Guasch-Ferre et al., 2016; Padberg et al., 2014). Glycosphingolipids are suggested to modulate β-cell signaling pathways implicated in diabetic disease such as apoptosis, β-cells cytokine secretion, islet autoimmunity and insulin gene expression (Boslem et al., 2012; Janikiewicz et al., 2015; Aerts et al., 2011). In a targeted metabolomics study, sphingomyelin was found to be negatively associated with risk of type 2 diabetes (Floegel et al., 2013), which is consistent with our findings in the integrated analysis showing decreased plasma levels of sphingomyelin in the subgroups at high risk of type 2 diabetes. A representative metabolite in de novo lipogenesis, palmitic acid, can mediate insulin signaling pathway and cause insulin resistance through increase synthesis of deleterious complex lipids and impaired function of cellular organelles (Palomer et al., 2018; Ma et al., 2015). Linoleic acid and its oxidized product HPODE, play an important role in inflammation, while decreased linoleic acid is suggested to predict insulin resistance and diabetes risk (Guasch-Ferre et al., 2016; Padberg et al., 2014; Roberts et al., 2014).

In the current study, PFASs exposure was strongly associated with alterations in numerous amino acid metabolism, which is generally consistent with the concept highlighted in a most recent review paper suggesting altered circulating levels of amino acids may modulate the risk of complications related to diabetes (Kahl and Roden, 2018). Specifically, we reported that tyrosine metabolism was one of the most affected pathways associated with PFASs exposure. Increased aromatic amino acids, such as tyrosine and phenylalanine, has been consistently found to be closely associated with hyperglycemia, insulin resistance and risk of developing diabetes in adults (Guasch-Ferre et al., 2016; Wang et al., 2011; Würtz et al., 2013), although the mechanisms underlying these associations are not fully understood. Muscle cells cultured with a mixture of amino acids including aromatic (e.g. tyrosine) and branched chain amino acids (BCAAs) resulted in activation of the
Fig. 3. Metabolic pathways associated with plasma PFASs concentrations. The vertical axis represents the pathways (blue circles) with circle radius representing the numbers of associated metabolite features, and horizontal axis represents the negative log$_{10}$ (p-value) of each pathway. (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

Fig. 4. Structural integrated analysis of PFASs plasma concentrations and individual metabolites for the identification of a subgroup of children with increased risk for developing type 2 diabetes. The thick blue line connecting “PFASs exposure” to “Cluster 2” indicates positive association, while thin blue line serves as the reference group. The blue lines connecting “Cluster” to metabolites suggest positive associations and grey lines suggest negative associations. The red line connecting “Cluster 2” and “Increased change in 2-hr glucose levels (OR=8.63)” represents that children in the latent “Cluster 2” were at higher risk for developing type 2 diabetes (OR = 8.63), compared to those in “Cluster 1” (reference). (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)
mammalian target of rapamycin (mTOR), impairment in insulin-stimulated phosphorylation of Akt/protein kinase B, and subsequently reduced glucose uptake (Tremblay and Marette, 2001). Although we did not find associations between PFASs exposure and alterations in the branched chain amino acids (BCAA) pathways, tyrosine can affect BCAA levels, since BCAA and amino acids compete for the same neutral amino acid transporter for cellular uptake (Adams, 2011; Fernstrom, 2005). We also found increased PFAS exposure was associated with the latent subgroup of participants at increased risk for developing type 2 diabetes, characterized by increased plasma arginine and decreased aspartate levels. These findings are in line with previous literature suggesting a positive association between arginine and risk of type 2 diabetes (Guasch-Ferres et al., 2016) and decreased aspartate and asparagine levels in adults developing type 2 diabetes (Palmer et al., 2015).

This study is strengthened by its longitudinal study design, the use of robust repeated clinical measures of insulin and glucose metabolism, and the use of novel statistical approaches to predict subgroups of Hispanic youth with increased susceptibility to type 2 diabetes based on their PFASs exposure and metabolomics profile. Detailed information regarding body fat percent and pubertal stage were also available and adjusted for in statistical models. Although we have adjusted for several factors that may affect glucose homeostasis, the latter is dependent on multiple exposures; thus, we cannot exclude the possibility of confounding from unmeasured variables such as diet. This proof-of-concept study was designed in a relatively-small sample size, however, we showed markedly changes in glucose homeostasis and alterations in metabolic pathways associated with PFASs exposure. Although it would be informative to examine differences by weight status, the current study is limited for this type of analysis as it is focused primarily on obese children. We found that PFASs exposure was only associated with change in 2-hour glucose levels. These findings indicate that PFASs exposure may negatively affect β-cell function. However, effects on insulin resistance cannot be ruled out. Future studies should examine the effects of PFASs exposure on type 2 diabetes using robust methods to assess changes in insulin resistance and β-cell function (e.g., glucose clamp techniques and intravenous glucose challenge tests) (Bergman et al., 1979; DeFronzo et al., 1979). The study focused on Hispanic children and findings are of public health significance as Hispanics have a disproportionate burden of environmental exposures, high rates of type 2 diabetes in youth, and are underrepresented in clinical research (Lascar et al., 2018; Mayer-Davis et al., 2017).

5. Conclusions

In summary, this is the first prospective study to demonstrate that PFASs exposure was associated with longitudinal alterations in glucose homeostasis in overweight and obese Hispanic youth. Dysregulation of several lipid and amino acid pathways that have been linked with type 2 diabetes were also associated with PFASs exposure. Lastly, a novel integrated latent variable analysis demonstrated that the observed changes in glucose homeostasis were characterized by increased PFASs levels and altered plasma metabolite profiles. Findings from this proof-of-concept study suggest that PFASs may play an important role in the pathogenesis of type 2 diabetes; however, larger studies are needed to replicate findings and to fully elucidate the mechanisms explaining the diabetogenic effects of PFASs exposure.

Authors’ contributions

Author contributions were as follows: conception and design of the study (LC, TLA, MIG); acquisition of data (DIW, TLA); analysis and interpretation of data (TLA, DIW, RJ, DV, FDG, MIG, KB, ZC, DVC, LC); drafting the article (TLA, JR, DIW, LC); and critical revisions for important intellectual content (TLA, DV, KB, DVC, DPJ, ZC, FDG, MIG, LC). LC had primary responsibility for final content. All authors read and approved the final manuscript.

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Conflict of interest

The authors declare that they have no conflict of interest.

Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi.org/10.1016/j.envint.2019.02.047.

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